

Hello. My name is Chuck Cravotta. I am a research hydrologist at the U.S. Geological Survey Pennsylvania Water Science Center. Today, I am going to introduce new software that I developed: "PHREEQ-N-AMDTreat+REYs" Water-Quality Modeling Tools to Evaluate Acid Mine Drainage Treatment Strategies for Recovery of Rare-Earth Elements.

Many studies have identified elevated concentrations of rare-earth elements plus yttrium (REYs) in acid mine drainage (AMD) and in solids formed in the treatment of these waters. Routine treatment methods plus additional steps may be applicable for the efficient recovery of REYs, depending on the AMD composition, volume, location, and economic considerations. The PHREEQ-N-AMDTreat+REYs geochemical modeling tools have the fundamental capability to simulate and predict key reactions during the routine treatment of AMD and the formation of treatment solids, including the precipitation of REYs compounds and the adsorption of REYs onto hydrous Fe, Al, and Mn oxides. The new tools were expanded from the PHREEQ-N-AMDTreat tools to include trace-element attenuation. Additions include a caustic titration model that indicates equilibrium surface and aqueous speciation of REYs as functions of pH and caustic agents, and a sequential kinetics+adsorption model that simulates progressive changes in pH, major ions, and REYs in water and solids during passive and/or active treatment. A goal of such modeling is to identify strategies that could produce a concentrated REYs extract from AMD or mine waste leachate that could be highly valuable. For example, if REYs could be concentrated after first removing substantial Fe and AI, the final REYs-bearing phase(s) could be more efficiently processed for REYs recovery and, therefore, may represent a more valuable commodity. Preliminary modeling supports the hypothesis that Fe and Al can be removed at pH >5.5 using conventional sequential oxidation and neutralization treatment processes without removing REYs, and that further increasing pH can promote the adsorption of REYs by hydrous Mn oxides. Alternatively, chemicals such as oxalate or phosphate may be added to precipitate REYs compounds following steps to decrease Fe and Al concentrations. Field and lab studies may be helpful to demonstrate REYs attenuation from AMD during specific treatment steps and to corroborate, refine, and constrain modeling parameters.

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AMDTreat 6.0 (2022) is a newly updated computer application for estimating costs and sizing of facilities to abate AMD (acidic or alkaline mine drainage) that is maintained by the Office of Surface Mining Reclamation and Enforcement (OSMRE).

The PHREEQ-N-AMDTreat water-quality modeling tool, developed by the USGS with support from OSMRE, was recently incorporated with the newly recoded AMDTreat 6.0 (beta version shown here).



The embedded PHREEQ-N-AMDTreat tool and the stand-alone TreatTrainMix2 tool reported by Cravotta (2020, 2021) simulate effects on water quality by various treatment components, involving different technologies and retention times. The model results may be used to optimize treatment system configuration or size and, also, for costs/benefits analysis.

Kinetics processes such as CO2 outgassing and O2 ingassing, Fell and MnII oxidation, limestone dissolution, and oxidation of organic matter are simulated, all of which may affect the pH and equilibrium speciation.

Active treatment with hydrogen peroxide and caustic chemicals also may be simulated. The potential for formation of solids from dissolved constituents and the corresponding mass of solids is computed, including amounts of Fe, Mn, and Al adsorbed by hydrous metal oxides (HMeO = HFO + HMO + HAO).

A newly expanded stand-alone model includes rare-earth elements attenuation by adsorption and precipitation.



The elements included in the PHREEQ-N-AMDTreat+REYs models are shown here in their respective positions on the periodic table. The lanthanide rare-earth elements plus yttrium and scandium are referred to as "REYs". REYs are in great demand for clean energy and other modern technologies.



The pH of coal-mine drainage in Pennsylvania has a bimodal frequency distribution. Roughly half of the discharges have pH less than 5.5. Dissolved concentrations of rareearth elements and other trace metals generally decrease with increased pH, with a break in slope at pH 5.5. Cravotta (2008) showed that trace-element minerals tend to be undersaturated in AMD and suggested adsorption as the primary mechanism to explain the relations among trace elements and pH.

Thus, the PHREEQ-N-AMDTreat+REYs tool set, which consists of the TreatTrainMix2REYS and CausticTitrationMix2REYs models, emphasizes the adsorption of REYs and other trace elements by hydrous metal oxides.



Rare-earth elements accumulate with manganese, aluminum, and iron that precipitate with other major elements in AMD treatment solids. The highest rare-earth element concentrations generally occur in solids with mixed compositions where manganese and aluminum are abundant. Such solids are typically produced by treatment of low-pH AMD. Solids dominated by iron can be produced by treatment of net-alkaline, near-neutral pH AMD that has relatively low dissolved concentrations of the rare earths to begin with.



At low pH, hydrous metal oxides tend to have positively charged surfaces that adsorb anions. As pH increases past the "zero point of charge", the surfaces become negatively charged and then adsorb cations.

In the PHREEQ-N-AMDTreat model, more than one compound may precipitate to form fresh sorbent: HMO consists of oxidized and reduced manganese hydroxide; HFO consists of oxidized and reduced iron hydroxide; and HAO consists of aluminum hydroxide and basaluminite.

In the PHREEQ-N-AMDTreat model, more than one compound may precipitate to form fresh sorbent: HMO consists of MnOOH (manganite) and Mn(OH)2 (pyrochroite); HFO consists of Fe(OH)3 (ferrihydrite) and Fe(OH)2; and HAO consists of Al(OH)3 and basaluminite.



This slide shows an example simulation of the adsorption of REYs (10⁻⁶ M each) as a function of pH, given a specified total mass of HMeO consisting of 0.03 g each of HFO, HMO, and HAO. The area below curve indicates the fraction that is adsorbed. Except for Sc and Ce, individual REYs have similar sorption behavior, with a majority sorbed by about pH 5.5, consistent with field observations.

Simulations were conducted using the CausticTitrationMix2REYsMoles.exe tool.

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Temp (C)	13.5	0.01	Cd (ug/L)	34.4	1E-08		1.925	1,91	4.6	<-Site density	sites/nm2		Aged	and/or					
SC (uS/cm)	5550	0	Co (ug/L)	4770	1E-08		0.2	0.0903	0.0405	<ste density<="" td=""><td>(weak or v) mol/</td><td>mal computed</td><td>freshly p</td><td>recipitated</td></ste>	(weak or v) mol/	mal computed	freshly p	recipitated					
DO (mg/L)	5.9	0.01	Cr (ug/L)	21.2	1E-08		0.005	0.1605]	<-Ste density	(strong or x) mol	mol computed	sorbent	properties					
pH	3	3	Cu (ug/L)	358	1E-08		HFO	HMO	HAO	FRES	HLY PRECIPITAT	ED SORBENT	(ADDITIONAL)						
Acidity (mg/L)	982	0	Ni (ug/L)	5110	1E-08		700	850	774	<surface are<="" td=""><td>a, m2/g 74809.7</td><td>90067.7 60</td><td>0376.3 Surface</td><td>area, m2/mol, comp</td></surface>	a, m2/g 74809.7	90067.7 60	0376.3 Surface	area, m2/mol, comp					
Estimate NetAcidity	1080.3	0	Pb (ug/L)	9.8	1E-08		1.925	1.91	4.6	<site density<="" td=""><td>sites/nm2</td><td></td><td>"Adjuste</td><td>d" model</td></site>	sites/nm2		"Adjuste	d" model					
Alk (mg/L)	0	0	Sc (ug/L)	149	1E-08		0.2334	0.1029	0.4613	<site density<="" td=""><td>(weak or y), mol/i</td><td>mal, computed</td><td>creased s</td><td>urface areas</td></site>	(weak or y), mol/i	mal, computed	creased s	urface areas					
TIC (mg/L as C)	19.2	0	Se (ug/L)	19.3	1E-08		0.0058	0.1829]	<site density<="" td=""><td>(strong or x), mol</td><td>mol, computed</td><td>of fresh p</td><td>recipitate.</td></site>	(strong or x), mol	mol, computed	of fresh p	recipitate.					
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Fe2 (mg/L)	29.6	0	Zn (ug/L)	18800	1E-08	51_54		10		SI Basalua	10	SI Mol	DHI2 0.0						
Estimate Fe2	0	0	La (ug/L)	201	1E-08	SI Fei	04)2	0.0		SI CaCO3	25	SI FeO	02 McCO3 25						
Al (mg/L)	128	1E-08	Ce (ug/L)	350	1E-08	5,_10	(SI	=0, equ	ilibriu	m precipi	tate; SI=99	, no preci	pitate)						
Mn (mg/L)	129	1E-08	Pr (ug/L)	66.4	1E-08		Specified	Saturation In	ndex Value	at Which Preci	pitation of REE W	II Occur-COMP	PETES WITH SC	RPTION					
SO4 (mg/L)	5000	1E-06	Nd (ug/L)	235	1E-08	SI_REE(OH)3	0.0	✓ SI_R	EE(CO3)1.5	5 0.0 ~	SI_REE(PO4)	0.0 、	SI_REE(C20	4)1.5 0.0					
Cl (mg/L)	1.9	0	Sm (ug/L)	79.7	1E-08	Select titrant:		Di	fferent	possible	caustic titr	ants		Maximum pH (<=1					
Ca (mg/L)	422	1E-06	Eu (ug/L)	23.1	1E-08	NaOH 6	~	wt% soln	⊖ Ca	(OH)2 (⊖ CaO	O Na2CO3	O CaCO3	11					
Mg (mg/L)	652	1E-06	Gd (ug/L)	99.3	1E-08	DUN N	0051		Select outp	ut matrix to be	saved and graphs	to display							
Na (mg/L)	17.8	0	Tb (ug/L)	21.3	1E-08	KUNIN	ODEL		O Short C	Output File	Long Out	tput File	Print PHRE	EQC Output Report					
K (mg/L)	3.46	0	Dy (ug/L)	122	1E-08	Plot	REYs_H	MeO	Plot R	EYs_ppt	Choose on-	screen gra	aphs						
Si (mg/L)	30.8	0	Ho (ug/L)	24.3	1E-08	Plot Sc	PI	XY 🖸	Plot La	Plot Ce	Plot Pr	Plot Nd	Plot Sm	Plot Eu					
NO3N (mg/L)	0.25	0	Er (ug/L)	67.4	1E-08	Plot Go	P	x Tb ⊡	Plot Dy	Plot Ho	Plot Er	Plot Tm	Plot Yb	Plot Lu					
PO4P (mg/L)	0.01	1E-11	Tm (ug/L)	8.85	1E-08	Plot	Cations_I	HMeO	Plot A	nions_HMeO	Plot Alkalini	ty 🗹 Plot Al	Plot Fe	Plot Mn					
F (mg/L)	0.5	0	Yb (ug/L)	54.4	1E-08	Plot Ca	Pl	x Mg	Plot Ba	Plot Sr	Plot Cd	Plot Co	Plot Cr	2 Plot 🖓					
DOC (mg/L as C)	2	0	Lu (ug/L)	7.82	1E-08	Plot Ni	PI	x Pb	Plot Zn	Plot U	Plot As	Plot Se	Plot PO	4 7 Not 504					
Oxalate (mg/L as C)	0.1	0	Y (ug/L)	600	1E-08	С	austicTitra	ionMix2RE	Ys.exe cre	ated by C.A. Cr	avotta III, U.S. Ge	ological Survey	. Release versio	n 1.0.0. June 2022					

This slide shows the user interface (UI) for the "CausticTitrationMix2REYs" tool. Redox and speciation reactions are modeled as instantaneous, equilibrium processes in response to changes in pH and sorbent availability. The model simulates the addition of a selected caustic agent (NaOH, Ca(OH)₂, CaO, Na₂CO₃, or CaCO₃) and the consequent attenuation of REYs, Fe, Mn, Al, and other elements by precipitation as solid compounds *and* by adsorption to hydrous metal oxide (HFO, HMO, and HAO).

The hydrous metal oxide sorbent can be a combination of fresh precipitate that forms during titration plus specified pre-existing solids. The freshly precipitated and existing sorbent can have different surface properties.

This particular image shows input values for untreated AMD at the Nittanny mine, which had low pH (3.0) with elevated concentrations of major and trace elements. The Nittanny AMD was titrated in the field with NaOH to pH 6.0, 7.5, 9.0, and 10.3. Measured concentrations of dissolved Fe, Al, Mn, and REYs decreased with increased pH.

Mg (652 mg L⁻¹), Ca (422 mg L⁻¹), Al (128 mg L⁻¹), Mn (129 mg L⁻¹), Fe (40.7 mg L⁻¹), and REYs (2.0 mg L⁻¹). Corresponding precipitated solids and sorbent properties were not characterized.



This slide shows graphs for the simulation results of the Nittanny AMD titration with NaOH, shown as curves, compared to measured values, shown as point symbols. Measurement errors are roughly the size of symbols.

The upper set of six graphs indicates results for default values of the specific surface area (Asp) of the "fresh" HFO, HMO, and HAO; REYs minerals were not allowed to precipitate. Decreases in the concentrations of Fe, Al, and Mn correspond to their precipitation as HFO, HAO, and HMO. The REYs begin to sorb at pH ~5 and are effectively sequestered to the freshly precipitated solids by pH 7.5. The individual graphs for La, Gd, and Lu indicate differences in the sorption affinities for different elements. HFO was the predominant sorbent of each of the REYs to about pH 8.5. Adsorption by HMO and HAO is indicated to be more important for La, which is considered a light REY, compared to Gd and Lu.

The lower set of graphs shows results after increasing the specific surface area of all three sorbents, especially that of HAO. The lower set also allows the precipitation of REYs compounds. For this scenario, adsorption by HMO and HAO are enhanced compared to the default simulation. Furthermore, precipitation of La and Gd as phosphate and carbonate compounds competes with sorption.

Both simulations provide reasonable fits to the empirical data in the pH range 3 to 9, with some improvement at high pH by the second simulation with greater surface area and REY precipitation.

These graphs were created using the selected output file imported to Excel.

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Select Workspace	C:\Users\cri	avotta\Documents	s \AMD Treat Train	REYs_wateq\	PineFor151212,	REYs					Knetics C	onstants, A	dust	ment Factors			_		Sorber	t Propertie	s, Specified HMeO + Equilibrium P		
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Acidty (mo/L)	0	0	Ni (m/l)	42	15-06		SL	Schwertmannit		1.0 ~	SI_Basal	minite	1	.0 ~	0 V SI_FeC03,MnC03			~	HAO, eq	ulppt	65 4.6		
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Estimate TIC	30.9	0	Sr (ug/L)	370	1E-06		O NaO	н 20 ч	~ wt%	soln C	Ca(OH)2	0	CaO	0	Na2CO3	CaO	03 (not kir	netic reacta	rt)	H2	02 wt% units gal/gal (memo, not us		
Fe (mg/L)	14	1E-08	U (ug/L)	0.03	1E-06	Segu	ential Treatme	nt Steps / Kinet	tics Con	dtions:	Aeration	Rate		Limestone	and Orga	nic Matter	Specifi	ed HMeO S	iorbent Co	ncentration	1		
Fe2 (mg/L)	14	0	Zn (ug/L)	37	1E-06	Step	+Caustic?->	pH? Time.hrs	Temp	2.C H202.mo	kLaC02.1/	s Lg(PCO)	2.atm)	SAcc.cm2/i	mol M/M	loc SOC.mol	HMeO	.mg Fe%	Mn%	Al%	Description		
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Al (mg/L)	0.09	1E-08	Ce (ug/L)	1.6	1E-06	₩ 2	1.5	0.0083	11.6		0.02	-3.4	-	33		0	1	197.8	10	12	Aeration nprap		
Mn (mg/L)	3.1	1E-08	Pr (ug/L)	0.18	1E-06	⊠ 3	1/5	13	12.10		0.00002	1.0.4	-	0			3	30	10		Uxidation/setting pond		
SO4 (mg/L)	330	1E-06	Nd (ug/L)	0.72	1E-06	⊻ 4	075	0.0028	12.16		0.005	-3,4	Ť	0		0	1	20	5	0	Aeraton cascade		
CI (mg/L)	4	0	Sm (ug/L)	0.11	1E-06		L 1/3	8	121		0.00005	-3.4	4	0		0.1	3	60	40		Aerobic weband		
Ca (mg/L)	56	1E-06	Eu (ug/L)	0.038	1E-06	⊠ e		0.0028	12.15		0.005	-3.4	-	0	10	0	1	60	40	10	Aeraton nprap		
Mg (mg/L)	51	1E-06	Gd (ug/L)	0.21	1E-06	27		6.1	12.04	0	0.00005	-3.4	-	0		0.1	2	40	60	0	Aerobic wetland		
Na (mg/L)	7.4	1E-06	Tb (ug/L)	0.018	1E-06	28		0.0028	12.04	0	0.005	-3.4	_	0	1	0	1	40	60	0	Aeration riprap		
K (mg/L)	0.54	0	Dy (ug/L)	0.073	1E-06	9		1.1	11.88	0	0.00001	3.4	~	0		0.1	2	20	03	0	Aerobic wetland		
Si (mg/L)	5.4	0	Ho (ug/L)	0.015	1E-06		0.	0.0042	11.88	0	0.005	-3.4	~	0		0	-	20	180	0	Aeration riprap		
NO3N (mg/L)	1.5	0	Er (ug/L)	0.024	1E-06	Ø 1	1:	0	11.88	0	10	-3.4	~	0	1	0	đ	100	10	1[0	TINULL		
PO4P (mg/L)	0.03	1E-11	Tm (ug/L)	0.0025	1E-06					Generate Sequential Kinetics Output										Print	PHREEQC Output Report		
F (mg/L)	0.1	0	Yb (ug/L)	0.015	1E-06		Plot De	Fe, Mn, A, D	0, NO3	Plot Ca, Na, Alk, Acidity Plot Sat Index						x		Plot PPT	Solids				
DOC (mg/L as C)	3.67	0	Lu (ug/L)	0.45	1E-06		Plot As	Se Co Cu Ni Pi	b Zn	P P	ot REvitot La	Ce Pr Nd S	Sm	D Pi	ot Eu Gd	To Dy Ho		Plot Er T	im Yb Lu Y	Sc			
Oxalate (mg/L as C)	1.3	1E-11	Y (ug/L)	0.59	1E-06				372	1000				TreatTr	rainMx2R8	Ys.exe created	by C.A.C	Cravotta III,	U.S. Geok	ogical Sun	vey. Release version 1.0.0. June		
Pine	Fo	rest	Min	e (Dec		201	5):				F		- /1		is exercised	in c.c.	Jaroud III.	0.3. 060				

The TreatTrainMix2REYs.exe tool simulates water-quality changes through sequential steps in passive or active AMD treatment systems. Kinetics processes such as CO_2 outgassing, O_2 ingassing, Fe^{II} and Mn^{II} oxidation, SO₄ reduction, and limestone dissolution, all of which affect pH, are coupled with the same equilibrium speciation and precipitation reactions as the CausticTitrationMix2REYs tool. Unlike the titration tool, the pH may decrease or increase in response to dynamic, kinetically limited processes. A total of 11 treatment steps may be considered, with each having a specified reaction time, CO_2 outgassing rate, availability of limestone, organic matter, H_2O_2 , sorbent, and other variables. A target pH may be specified for the addition of a caustic agent (NaOH, Ca(OH)₂, CaO, Na₂CO₃, or CaCO₃) to begin steps 1 to 5, possibly after aeration (decarbonation) or other pre-treatment steps. The solution composition at the end of each step is passed to the next step.

This slide shows the UI with input values for simulation of passive treatment of net-acidic coal mine discharge at the Pine Forest Mine in Pennsylvania. Sequential treatment involves an anoxic limestone drain at step 1; aeration pond at step 3; and aerobic wetlands at steps 5, 7, and 9, with aeration at intermediate and final steps (2, 4, 6, 8, 10).

For the reported simulations, retention time, CO_2 outgassing rate, available limestone, and pre-existing HMeO mass and composition were varied at each step. The fractions of Fe, Al, and Mn in the pre-existing sorbent were estimated from measured sediment composition at the inflow and points downstream of the ALD.

HMeO mass of 75 mg L⁻¹ specified for the ALD represents accumulated coatings on

limestone particles. For subsequent steps, the specified sorbent mass was only 1 to 3 mg, representing suspended particles or coatings on rock or plant surfaces. Consistent values for retention time, kinetics variables, and sorbent properties were used for different model scenarios whereby the only values varied were the rate factor for iron-oxidizing bacteria (FeOB) and/or potential for REYs solids to precipitate. For the simulated "biofouling" scenario, the FeOB rate factor (factr.kbact) was increased from the default value of 1 to 2; for the abiotic scenario, that factor is 0.

HMeO mass of 75 mg L⁻¹ specified for the ALD is consistent with a 0.22-mm thick coating on limestone particles (72 cm² mol⁻¹) in contact with 1 L water volume, assuming 35 % bed porosity and sorbent density of 1.92 g cm⁻³.



This Google Earth image from 2012 shows sampling points along flow path through the Pine Forest passive treatment system. The first step is an underground "biofouled" anoxic limestone drain that had gelatinous orange precipitate coating limestone particles and clogging porosity. Water started to upwell to the surface along the length of the bed because of accumulated solids.

A settling tank near the inflow permitted initially anoxic AMD to become partly oxygenated before entering the buried limestone bed.

Subsequent treatment steps involve aerobic ponds and wetlands with aeration steps in between. Total retention time in the system is approximately 30-40 hours, depending on inflow rate.



The TreatTrainMix2REYs results shown as a function of retention time generally reproduce the longitudinal trends for measured water quality at the Pine Forest treatment system. The pH and DO increased within the ALD and, subsequently, in response to aeration steps. The Fe concentration decreased by 30% within the ALD, simulated to result from abiotic plus microbial oxidation of Fe^{II}, which led to the precipitation and accumulation of HFO with lesser quantities of HAO and HMO on limestone surfaces. Concentrations of Mn were relatively unchanged through the ALD but decreased by about 30% within wetlands where the pH and DO were greatest and the Mn content of accumulated sorbent was highest. Initial concentrations of Al were very low and were simulated to decrease by adsorption to accumulated sorbent within the ALD. Simulated attenuation of REYs also took place within the ALD because of adsorption. Despite less mass of HMeO within wetlands downstream of the ALD, greater Mn content of sorbent and increased pH in wetlands promoted attenuation of dissolved Mn^{II} and remaining REYs.

Compared to adsorption, REYs precipitation has only a small effect as indicated by the difference between red-dashed (REY precipitation) and blue-dashed curves (no REY precipitation), both simulating the biofouling scenario where the FeOB rate factor was doubled. Simulation results for the two reference scenarios (orange solid or black dotted curves) demonstrate that abiotic, homogeneous Fe^{II} oxidation may explain observations in aerobic ponds and wetlands, but does not explain observed Fe removal within and immediately below the ALD. Thus, microbial and abiotic oxidation, combined with

adsorption account for the attenuation of Fe, REYs, and associated metals within the limestone bed. However, in downstream zones where pH and DO are greater, abiotic Fe^{II} oxidation is sufficient to explain observed trends.



The TreatTrainMix2REYs tool is also used to simulate decreasing concentrations of dissolved metals and REYs within the Silver Creek passive treatment system. This Google Earth image from 2012 shows sampling points along conceptual flow path through the passive treatment system, which consists of a sedimentation pond at step 1, two large oxidation/settling ponds at steps 3 and 5, followed by two aerobic wetlands at steps 7 and 9, with wide, shallow aeration cascades in between.

The second and third ponds account for 4/5 of the total retention time for the system, which is approximately 250-450 hours, depending on inflow rate.

When sampled in 2015 and 2016, the Silver Creek AMD was anoxic, marginally net alkaline with pH 5.9-6.0. Rapid outgassing of CO_2 during aeration steps caused large increases in pH, which facilitated Fe^{II} oxidation in subsequent steps.



For initial simulations, HMeO composition was varied on the basis of values reported by Ashby (2017) for sampled sediments at Silver Creek. The HMeO mass at each step was assumed to be less than or equal to the difference between immediately upstream and downstream samples for the combined mass of Fe, Al, and Mn. Only the CO_2 outgassing rate and sorbent mass and composition (HMeO.mg, Fe%, Mn%, Al%) at each step were adjusted to achieve a reasonable match between empirical and simulated values for longitudinal changes in pH, Fe, Mn, Al, and associated major solute concentrations. Eventual removal of Mn^{II} in the wetland treatment steps were simulated by HMeO sorbent having greater HMO content, as observed for the sampled sediment.



Abiotic oxidation of Fe combined with adsorption and precipitation of solids explains observed attenuation of Fe, Al, Mn, and associated REYs at the Silver Creek treatment system; microbial Fe oxidation had little effect (FeOB=1X). Although the results for initial simulations effectively reproduced the longitudinal trends for measured pH, DO, Fe, Mn, and Al (black dashed or orange curves), without REYs mineral precipitation, the corresponding modeled concentrations of total dissolved REYs (and individual REYs, not shown) remaining in solution were at least five times greater than observed values for all but the last stages of the treatment system (black dashed curves). Simply permitting the precipitation of REYs (SI_REEPO4=0) resulted in substantial decrease in the concentration of total REYs (blue or red curves), consistent with observations. Nevertheless, simulated attenuation was not consistently effective for individual REYs, many of which remained undersaturated (e.g. Y, Eu, Gd, Tb, Dy, Ho, Yb, and Lu). As explained by Liu and Byrne (1997), formation of REEPO₄ in the environment generally involves co-precipitation of phases which incorporate multiple REYs, in which case the overall trend may be expected.



A goal of modeling with the PHREEQ-N-AMDTreat+REYs tools is to identify strategies that could feasibly produce a concentrated REYs extract from AMD or mine waste leachate that could be highly valuable.

For example, leachate from pyritic shale and coal waste (coal refuse) at a centralized processing facility has low pH and elevated concentrations of metals. The acidic, metalladen leachate is a long-term treatment liability that will persist long after coal mines have closed. An economically sustainable approach for recovery of REYs from the leachate could offset treatment costs.

If Fe and Al can be removed sequentially, at pH < 5.5, REYs may be retained in solution and recovered at pH > 6.5 with manganese in later steps.



Untreated leachate from a coal-refuse disposal facility in Pennsylvania is considered as a proposed test case for REYs recovery. Current treatment utilizes neutralization with lime, which causes precipitation of Fe, Al, and associated REYs into a complex sludge mixture. If REYs could be concentrated after first removing substantial Fe and Al, the final REYs-bearing phase(s) may be efficiently processed for REYs recovery.

The TreatTrainMix2REYs tool was used to simulate the current active treatment with lime to pH 8.5 and two alternative treatments intended to concentrate REYs from the leachate.

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Na (mg/L)	83.2	1E-06	Tb (ug/L)	17.1	1E-06		8:	0	25	0	0	-3.4	~)	1	0	0	0	0	0	NULL	
K (mg/L)	18.8	0	Dy (ug/L)	95.6	1E-06		9:	0	25	0	0	-3.4	~)	1	0	0	0	0	0	NULL	
Si (mg/L)	25.9	0	He (ug/L)	18.5	1E-06		10:	0	25	0	0	-3.4	~)	1	0	0	0	0	0	NULL	
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This slide shows the user interface for the lime treatment scenario.



This slide shows the user interface for the alternative treatment scenario using H2O2 followed by NaOH addition and settling and aeration steps.

Numerican and account and account and account a	Coloret Wedges and	Callerate			05%	0001-13-00	¥-	Tre	atTra	ain	Mix2	REYs	.e	æ									
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DD PurUl Emblem Content of Statement of Sta	SC (uS/am)	9740	0	Co (ug/L)	669	1E-06		fact	r kSHFO		1	factr kSO	с	100	_	factr k	DOC	0.1		HMO e	autort	746	191
H 12 0 Children 0 State State <th< td=""><td>DO (mg/L)</td><td>6.07</td><td>0.01</td><td>Cr (ug/L)</td><td>77.5</td><td>1E-06</td><td></td><td></td><td></td><td></td><td>Eaulibr</td><td>um Constants</td><td>Adustr</td><td>ent of Sat</td><td>uration Inde</td><td>x for Precis</td><td>otation</td><td></td><td></td><td>HAO, H</td><td>MeO</td><td>68</td><td>4.6</td></th<>	DO (mg/L)	6.07	0.01	Cr (ug/L)	77.5	1E-06					Eaulibr	um Constants	Adustr	ent of Sat	uration Inde	x for Precis	otation			HAO, H	MeO	68	4.6
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NOXIM pol 0.01 0 Er kg/L 68.7 (E6.6) File (C) 0 0 0 0 0 0 NULL POLP inp(L) 2.0 (E6.1) Tin kg/L (E6.7) (E6.6) Generation Sequental Kristics Output	Si (mg/L)	25.9	0	Ho (ug/L)	18.5	1E-06		10:	0.033	25	10	10.005	-3.4			1	0	0.5	10	85	5	Ditch	
PROP PropU 20 [E11] The log U 657 [E46] [Demende Sequenti Mentol Odual Image: PropU PropU 05 0 PropU 05.5 [E46] [Demende Sequenti Mentol Odual Image: PropU PropU 05.5 [E46] [Demende Sequenti Mentol Odual Image: PropU PropU PropU 0.5 [Demende Sequenti Mentol Odual Image: PropU PropU PropU 0.5 [Demende Sequenti Mentol Odual Image: PropU PropU PropU 0.5 [Demende Sequenti Mentol Odual [PropU PropU PropU <td>NO3N (mg/L)</td> <td>0.01</td> <td>0</td> <td>Er (ug/L)</td> <td>48.7</td> <td>1E-06</td> <td></td> <td>11:</td> <td>0</td> <td>25</td> <td>0</td> <td>0</td> <td>-3.4</td> <td>~ [</td> <td></td> <td>1</td> <td>0</td> <td>0</td> <td>0</td> <td>0</td> <td>0</td> <td>NULL</td> <td></td>	NO3N (mg/L)	0.01	0	Er (ug/L)	48.7	1E-06		11:	0	25	0	0	-3.4	~ [1	0	0	0	0	0	NULL	
Filip(1) 0.5 0 Yb: bg/L 55.4 [E66 CP Rd: Dia Fe, Min, AL 00, NO3 P Rd: Ex, No., Mik, Addaty P Rd: Ex Index P Rd: TPT 564da 0000 (mp/L as C) 0.5 0 LLIS/L 1E665 P Rd: Re, Min, AL, Ook, NO3 P Rd: Ex Index P Rd: Tim Yb, LUT Sc Dublete (mp/L as C) 0.1 1E611 Y lisp/L 4421 1E665 Transfit and Miz/REY's are constrained by CA: Carents III, U.S. Geological Survey, Release version 1.00. J	PO4P (mg/L)	2.0	1E-11	Tm (ug/L)	6.07	1E-06						G	ienerate	Sequentia	Kinetics 0	tput					Prin	PHREEQC Out	tput Report
000 Peg Lan C 0.5 0 Lu Leg L 523 1E-66 Pet As Se Co Cu N Pb Zn Pet Ar Se Co Cu N Pb Zn Pet As Se Co Cu	F (mg/L)	0.5	0	Yb (ug/L)	35.4	1E-06		Plot Dis	Fe, Mn, Al, D	0, NO3	D Pi	t Ca, Na, Ak,	Acidity		Plo	Sat Index			Plot PPT	T Solids			
Dealer ImpLan D 0.1 1E11 Y lop/U 481 1E06 TestTanMo2RDYneer createdby CA Drivita III, US Gelogiad Surviv, Release venion 10.0. J	DOC (mg/L as C)	0.5	0	Lu (ug/L)	5.29	1E-06		Plot As	Se Co Cu Ni P	b Zo	D Pk	REYtot La	Ce Pr Nd	Sm		Eu Gd Ta	Dy Ho		Plot Er	Tm Yb Lu 1	Sc		
	Oxalate (mg/L as C)	0.1	1E-11	Y (ug/L)	481	1E-06]				1.000				Treat Tra	inMx2RE)	is exe create	dby CA O	ravotta III	U.S. Geol	ogical Sur	vev. Belease w	ension 1.0.0. June
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This slide shows the user interface for the alternative treatment scenario using H2O2 followed by Na2CO3 addition and settling and aeration steps.



The active lime treatment to pH 8.5 results in nearly complete removal of REYs with Fe-Al-Ca rich sludge. REYs, which are adsorbed primarily by HFO and HAO, are diluted by the major metals and other impurities in the sludge.

By comparison, alternative treatment strategies using H_2O_2 to oxidize Fe^{II}, followed by metered addition of Na-caustic agents (NaOH or Na₂CO₃) to achieve a target pH < 5.5 sequentially remove Fe and Al from solution. Subsequent aeration of the effluent over extended time results in the oxidation of Mn which adsorbs and concentrates REYs in final steps. The scenario using NaOH instead of Na2CO3 maintains pH < 5.5 until most Fe and Al have been removed and, consequently, results in greater concentrations of REYs with Mnrich solids.

Alternatively, chemicals such as oxalate or phosphate may be added to the effluent to precipitate REYs compounds, following initial steps to remove Fe and Al



The PHREEQ-N-AMDTreat+REYs models suggest that conventional treatment technology to recover REYs and other critical minerals may be feasible. Nevertheless, economic benefits warrant further evaluation. In this 2021 paper by Fritz and others, chemical costs plus additional factors such as purification and transportation costs were considered in their technoeconomic assessment of REYs recovery from AMD.



As suggested by Fritz et al., use of sorbents may be economically favorable over methods involving REYs precipitation. Thus, the alternative treatment strategies simulated previously, involving sorption by HMO, could be advantageous over other methods. The HMO sorbent may possibly be re-used after recovery of REYs. Additional research on specific treatment strategies and associated economic factors would be needed for any given site-specific design.

Conclusions

- Geochemical tools using PHREEQC that incorporate equilibrium aqueous and surface speciation *plus* kinetics models for oxidation processes and limestone dissolution have been developed to evaluate mine effluent treatment options.
- The user interface for the PHREEQ-N-AMDTreat tools facilitates adjustment to input concentrations and system variables.
- Graphical and tabular output indicates changes in the pH, solute concentrations, and element distributions in effluent and associated solids.
- ✓ By adjusting kinetics variables or chemical dosing, various passive and/or active treatment strategies can be simulated.
- AMDTreat cost-analysis software can be used to evaluate the feasibility for installation and operation of treatments that produce the desired effluent quality.

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Cravotta, C.A. III (2022) Interactive PHREEQ-N-AMDTreat+REYs water-quality modeling tools to evaluate potential attenuation of rare-earth elements and associated dissolved constituents by aqueous-solid equilibrium processes: U.S. Geological Survey Software Release (software download). https://doi.org/10.5066/P9M5QVK0

To use the executable models, IPhreeqcCOM for Windows (Charlton and Parkhurst, 2011) must be installed on the user's computer. That software is accessible for download at: https://water.usgs.gov/water-resources/software/PHREEQC/IPhreeqcCOM-3.7.3-15968-win32.msi

https://water.usgs.gov/water-resources/software/PHREEQC/IPhreeqcCOM-3.7.3-15968-x64.msi

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