USER-FRIENDLY GEOCHEMICAL MODELING TO EVALUATE PASSIVE AND ACTIVE TREATMENT FOR COAL-MINE DISCHARGES

Charles A. Cravotta III, Research Hydrologist, U.S. Geological Survey, Pennsylvania Water Science Center, 215 Limekiln Rd, New Cumberland, PA 17070

Extended Abstract:

User-friendly aqueous geochemical tools were developed to simulate changes in water quality resulting from passive and active treatment of coal-mine drainage. The interactive geochemical tools utilize PHREEQC equilibrium (aqueous and surface speciation) and kinetics models (Parkhurst and Appelo, 2013) for gas exchange; iron and manganese oxidation and precipitation; limestone dissolution; and organic carbon oxidation combined with reduction of nitrate, sulfate, and ferric iron. Reactions with synthetic caustic chemicals (CaO, Ca(OH)₂, NaOH, Na₂CO₃) or oxidizing agents (H₂O₂) also may be simulated separately or combined with sequential kinetic steps. A user interface facilitates input of initial water chemistry data and adjustment of kinetic variables. Graphical and tabular output indicates the changes in pH, metals and other solute concentrations, total dissolved solids, and specific conductance of treated effluent plus the cumulative quantity of precipitated solids as a function of detention time or the amount of caustic agent added. By adjusting kinetic variables or chemical dosing, the effects of independent or sequential treatment steps that have different detention time (volume/flow rate), aeration rate, quantities of reactive solids, and temperature can be simulated for a specified influent quality. Thus, various passive and/or active treatment strategies can be identified to achieve the desired effluent quality. Then, AMDTreat (e.g. Cravotta et al., 2015) can be used to evaluate the potential cost-effectiveness of treatments considering available land area, equipment, and funding for construction and operation.

The kinetics tool for atmospheric exchange is coupled with those for iron and manganese oxidation. A first-order rate law describes CO₂ outgassing and O₂ ingassing, which affect the pH and rates of iron and manganese oxidation (e.g. Cravotta, 2015). Different aeration technologies may be considered for which gas-exchange rate constants are assigned. The iron and manganese oxidation kinetics tools utilize pH-dependent rate models that consider the concentrations of dissolved O₂ (DO) and Fe^{II} or Mn^{II} plus catalysis by surfaces (heterogeneous oxidation) and microbes (biotic oxidation) (Stumm and Lee, 1961; Tamura et al., 1976; Davies and Morgan, 1989; Vries et al., 2017). Microbial catalysis rates are adjusted for pH, DO,

temperature, and the concentration of iron oxidizing bacteria (FeOB). The acidophilic FeOB rate increases as pH decreases from 5 to 2.8 and exceeds the abiotic Fe^{II} oxidation rate at low pH (e.g. Kirby et al., 1999), whereas the neutrophilic FeOB rate generally does not exceed the abiotic rate, except at optimum conditions of near-neutral pH (6.5-7.5) and low DO (1.9-2.2 mg/L), where the FeOB rate can be 20 times the abiotic heterogeneous rate (Eggerichs et al., 2014). The Arrhenius equation is used to adjust the Fe^{II} and Mn^{II} oxidation rate constants from the reference temperature of 20 °C to other temperatures (Appelo and Postma, 2005).

The heterogeneous oxidation rate models incorporate pH-dependent adsorption of Fe^{II} and Mn^{II} by hydrous ferric oxide (HFO) (Dzombak and Morel, 1990) and Mn^{II} by hydrous manganese oxide (HMO) (Tonkin et al., 2004). Increasing the available mass (surface area) of HFO or HMO, for example the recirculation of HFO solids within a reaction tank or the accumulation of HMO on limestone particles in a manganese removal bed, increases potential for attenuation of dissolved metal concentrations at a given pH.

The limestone kinetics tool simulates dissolution (and precipitation) of calcite in passive systems, including anoxic or oxic limestone beds, manganese removal beds, vertical flow pond (VFP), and compost mixes. The rate expression considers solution chemistry, mainly pH and partial pressure of CO₂, plus the surface area and purity of limestone particles (Plummer et al., 1978; Appelo et al., 1998). Limestone particle surface area and volume are estimated assuming an ellipsoid shape for various aggregate sizes (e.g. Cravotta et al., 2008; Pennsylvania Department of Environmental Protection, 2000).

Reactions within a compost layer of a VFP or a biochemical reactor are simulated by the dissolution of limestone fines combined with the oxidation of organic carbon by DO, nitrate, and sulfate. A user specifies the volumetric fractions of organic material and limestone and the size of limestone particles. Aqueous sulfide produced from the reduction of sulfate by organic carbon will precipitate as FeS or adsorb to HFO. To simulate the reduction of Fe^{III} in the compost layer, a kinetics expression couples the rates of oxidation of the adsorbed sulfide and the reductive dissolution of HFO (e.g. dos Santos Alfonso and Stumm, 1992; Peiffer et al., 1992; Poulton, 2003).

The presentation demonstrates the model to evaluate changes in water quality that may result from commonly used treatment steps for acidic or alkaline influent and describes associated implications for treatment system design and optimization. Figures 1 and 2 give examples of user input and model output, and Table 1 shows possible sizing calculations.

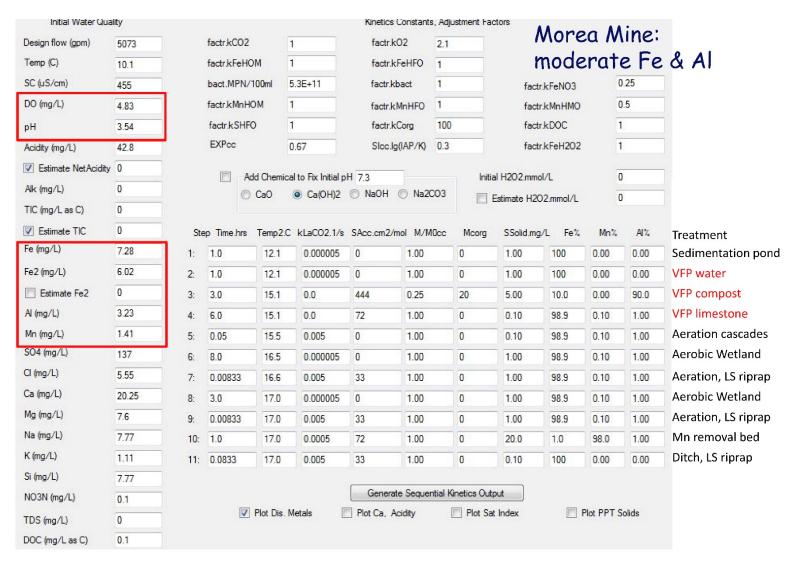


Figure 1: User interface (UI) for simulation of sequential passive treatment of net-acidic coal mine discharge through (1) sedimentation pond; (2-4) vertical flow pond (VFP); (6) aerobic pond; (8) aerobic wetlands; and (10) manganese removal bed with intermediate aeration steps (5, 7, 9, 11). Water-quality results of treatment are shown in Figure 2. Detention times are used to estimate approximate size of treatment system in Table 1.

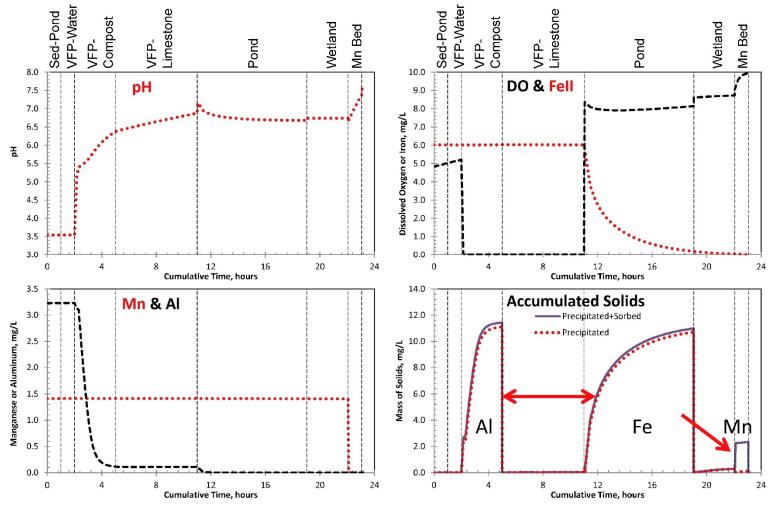


Figure 2: Results of PHREEQC kinetics model for passive treatment of coal-mine drainage as a function of the cumulative detention time for 11 numbered sequential steps (Fig. 1, Table 1): (1) 4-ft deep sediment pond; (2) 2-ft deep VFP water layer, (3) 2-ft thick VFP compost layer composed of 25% limestone fines and 75% organic matter having 45% porosity, (4) 4-ft thick VFP limestone layer having 45% porosity; (6) 4-ft deep aerobic pond; (8) 1-ft deep aerobic wetlands; and (10) 0.5-ft deep manganese removal bed, with aeration steps (5, 7, 9, 11) between major treatment steps.

Table 1. Example treatment system sizing calculations based on detention time estimates used in PHREEQC kinetics model

Step	Treatment	flow rate, ft3/s	deten- tion time, sec	deten- tion time, hr	depth,	porosity	volume, ft3	area of water surface, ft2	area of water surface, acres	lime- stone particle size, AASHTO	lime- stone particle size, cm2/mol	CaCO3 fraction in bulk, M/M0cc	lime- stone mass, tons	compost organics mass, tons
<u>Step</u> 1	Sedimentation pond	11.30	3600	1.00	4.00	1.00	40697	10174	0.23	AASITIO	CITIZ/IIIOI	1.00	0	_
	·							10174	0.23					0
2	VFP water	11.30	3600	1.00	2.00	1.00	40697					1.00	0	0
3	VFP compost	11.30	10800	3.00	2.00	0.45	271310			8	444	0.25	3086	6244
4	VFP limestone	11.30	21600	6.00	4.00	0.45	542620	106828	2.45	3	72	1.00	24686	0
5	Aeration	11.30	180	0.05	0.10	1.00	2035	20348	0.47			1.00	0	0
6	Oxidation Pond	11.30	28800	8.00	4.00	1.00	325572	81393	1.87			1.00	0	0
7	Aeration	11.30	30	0.01	0.10	0.45	753	7533	0.17	R-3	33	1.00	34	0
8	Aerobic Wetland	11.30	10800	3.00	1.00	1.00	122090	122090	2.80			1.00	0	0
9	Aeration	11.30	30	0.01	0.10	0.45	753	7533	0.17	R-3	33	1.00	34	0
10	Mn removal bed	11.30	3600	1.00	0.50	0.45	90437	180873	4.15	3	72	1.00	4114	0
11	Ditch	11.30	300	0.08	0.50	0.45	7533	15067	0.35	R-3	33	1.00	343	0
1 to	Tatal	·		00.45	40.00	·		554040	40.00		·		20000	0044
11	Total:			23.15	18.30			551840	12.66				32298	6244

Volume is computed as the product of flow rate and detention time. Design flow rate of 5073 gal/min = $11.30 \, \text{ft}^3/\text{s}$.

Area is computed as the volume divided by depth; for the VFP, volumes and depths for each of the three steps are summed before computing area. Masses of limestone and compost are computed as the product of their respective volume and bulk density.

Selected References:

- Appelo, C.A.J., and Postma, D., 2005, Geochemistry, groundwater and pollution (2nd). Balkema, Leiden, 678 p.
- Appelo, C.A.J., Verweij, E., and Schäfer, H., 1998. A hydrogeochemical transport model for an oxidation experiment with pyrite/calcite/exchangers/organic matter containing sand. Applied Geochemistry 13, 257–268.
- Ball, J.W., and Nordstrom, D.K., 1991. User's manual for WATEQ4F, with revised thermodynamic data base and test cases for calculating speciation of major, trace, and redox elements in natural waters. U.S. Geological Survey Open-File Report 91-183.
- Burrows, J.E., Cravotta, C.A. III, and Peters, S.C., 2017. Enhanced Al and Zn removal from coalmine drainage during rapid oxidation and precipitation of Fe oxides at near-neutral pH: Applied Geochemistry, v. 78, p. 194-210.
- Charlton, S.R., and Parkhurst, D.L., 2011. Modules based on the geochemical model PHREEQC for use in scripting and programming languages. Computers & Geosciences 37, 1653-1663.
- Cravotta, C.A. III, 2007. Passive aerobic treatment of net-alkaline, iron-laden drainage from a flooded underground anthracite mine, Pennsylvania, USA. Mine Water and the Environment 26, 128-149.
- Cravotta, C.A. III, 2015. Monitoring, field experiments, and geochemical modeling of Fe(II) oxidation kinetics in a stream dominated by net-alkaline coal-mine drainage, Pennsylvania, U.S.A. Applied Geochemistry 62, 96-107.
- Cravotta, C.A. III, Ward, S.J., and Hammarstrom, J.M., 2008. Downflow limestone beds for treatment of net-acidic, oxic, iron-laden drainage from a flooded anthracite mine, Pennsylvania, USA--Laboratory evaluation: Mine Water and the Environment 27, 86-99.
- Cravotta, C.A. III, Means, B., Arthur, W., McKenzie, R., and Parkhurst, D.L., 2015. AMDTreat 5.0+ with PHREEQC titration module to compute caustic chemical quantity, effluent quality, and sludge volume. Mine Water and the Environment 34, 136-152.
- Davies, S.H.R., and Morgan, J.J., 1989. Manganese(II) oxidation kinetics on metal oxide surfaces. Journal of Colloid Interface Science 129, 63-77.
- Davison, W., and Seed G., 1983. The kinetics of the oxidation of ferrous iron in synthetic and natural waters. Geochimica et Cosmochimica Acta 47, 67-79.
- Dempsey, B.A., Roscoe, H.C., Ames, R., Hedin, R., and Byong-Hun, J., 2001. Ferrous oxidation chemistry in passive abiotic systems for the treatment of mine drainage. Geochemistry Exploration, Environment, Analysis 1, 81-88.
- dos Santos Alfonso, M., and Stumm, W., 1992. Reductive dissolution of iron(III) (hydr)oxides by hydrogen sulfide. Langmuir 8, 1671-1675.
- Dzombak, D.A., and Morel, F.M.M., 1990. Surface complexation modeling: Hydrous ferric oxide. John Wiley and Sons, New York, NY, USA, 393 p.
- Eckert, P., Appelo, C.A.J., 2002. Hydrogeochemical modeling of enhanced benzene, toluene, ethylbenzene, xylene (BTEX) remediation with nitrate. Water Resources Research 38,1130, 5.1-5.11.
- Eggerichs, T., Opel, O., Otte, T., and Ruck, W., 2014. Interdependencies between biotic and abiotic ferrous iron oxidation and influence of pH, oxygen and ferric iron deposits. Geomicrobiology Journal 31, 461-472.
- Geroni, J.N., Sapsford, D.J., 2011. Kinetics of iron(II) oxidation determined in the field. Applied Geochemistry 26, 1452-1457.

- Geroni, J.N., Cravotta, C.A. III, Sapsford, D.J., 2012. Evolution of the chemistry of Fe bearing waters during CO₂ degassing. Applied Geochemistry 27, 2335-2347.
- Kirby, C.S., and Cravotta, C.A., III, 2005. Net alkalinity and net acidity 2 -- Practical considerations. Appl. Geochem. 20, 1941-1964.
- Kirby, C.S., Thomas, H.M., Southam, G., Donald, R., 1999. Relative contributions of abiotic and biological factors in Fe(II) oxidation in mine drainage. Applied Geochemistry 14, 511-530.
- Kirby, CS, Dennis, A, Kahler, A., 2009. Aeration to degas CO₂, increase pH, and increase iron oxidation rates for efficient treatment of net alkaline mine drainage. Applied Geochemistry 24, 1175-1184.
- Parkhurst, D.L., and Appelo, C.A.J., 2013. Description of input and examples for PHREEQC version 3—A computer program for speciation, batch-reaction, one-dimensional transport, and inverse geochemical calculations. USGS Techniques Methods 6-A43, 497 p.
- Peiffer, S., dos Santos Afonso, M., Wehrli, B., and Gachter, R., 1992. Kinetics and mechanism of the reaction of H₂S with lepidocrocite. Environmental Science & Technology 26, 2408-2412.
- Peiffer, S., 2016. Reaction time scales for sulphate reduction in sediments of acidic pit lakes and its relation to in-lake acidity neutralization. Applied Geochemistry 73, 8-12.
- Pennsylvania Department of Environmental Protection, 2000. Erosion and sediment pollution control program manual: Harrisburg, Pennsylvania Dept. Environmental Protection Bureau of Watershed Management, Document No. 363-2134-008, 180 p. (tables 9 and 10A).
- Pesic, B., Oliver, D.J., and Wichlacz, P., 1989. An electrochemical method of measuring the oxidation rate of ferrous to ferric iron with oxygen in the presence of *Thiobacillus ferrooxidans*. Biotechnology and Bioengineering 33, 428-439.
- Plummer, L.N., Wigley, M.L., and Parkhurst, D.L., 1978. The kinetics of calcite dissolution in CO₂-water systems at 5° to 60°C and 0.0 to 1.0 atm CO₂. American Journal of Science 278, 179-216.
- Poulton, S.W., 2003. Sulfide oxidation and iron dissolution kinetics during the reaction of dissolved sulfide with ferrihydrite. Chemical Geology 202, 79-94.
- Rathbun, R.E., 1998. Transport, behavior, and fate of volatile organic compounds in streams: US Geological Survey Professional Paper 1589, 151 p.
- Skousen J.G., Sextone, A., and Ziemkiewicz, P.F., 2000. Acid mine drainage control and treatment. In, Barnhisel, R.I., Daniels, W.L., Darmody, R.G. (Eds.) Reclamation of drastically disturbed lands. American Society of Agronomy Monograph 41, 131-168.
- Stumm, W., and Lee, G.F., 1961. Oxygenation of ferrous iron. Industrial and Engineering Chemistry 53, 143-146.
- Stumm, W., and Morgan, J.J., 1996. Aquatic chemistry--chemical equilibria and rates in natural waters (3rd). New York, Wiley-Interscience, 1022 p. (especially p. 682-691)
- Sung, W., and Morgan, J.J., 1980. Kinetics and product of ferrous iron oxygenation in aqueous systems. Environmental Science & Technology 14, 561-568.
- Tamura, H., Goto, K., and Nagayama, M., 1976. The effect of ferric hydroxide on the oxygenation of ferrous iron in neutral solutions. Corrosion Science 16, 197-207.
- Tonkin, J.W., Balistrieri, L.S., and Murray, J.W., 2004. Modeling sorption of divalent metal cations on hydrous manganese oxide using the diffuse double layer model. Applied Geochemistry 19, 29-53.
- Vries, D., Bertelkamp, C., Schoonenberg Kegel, F., Hofs, B., Dusseldorp, J., Bruins, J.H., de Vet, J.H., and van den Akker, B., 2017. Iron and manganese removal: Recent advances in modelling treatment efficiency by rapid sand filtration. Water Research 109, 35-45.